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Restoration Strategies Regional Water Quality Plan – Science Plan for the Everglades Stormwater Treatment Areas:

Soil Amendments/Management to Control P Flux

Phase I Summary Report for the Use of Soil Amendments/Management to Control P Flux Study

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Study Plan Summary

Anaerobic wetland soils mobilize forms of phosphorus (P) that are usually retained in welldrained terrestrial soils. As a result, P moves much more readily through wetland soils and can flux (i.e., diffuse) upward into the water column provided there is a decreasing concentration gradient between the soil porewater and the overlying water column (Reddy and DeLaune 2008, Kadlec and Wallace 2009, Ballantine and Tanner 2010). The objective of the Use of Soil Amendments/ Management Techniques to Control P Flux Study is to investigate whether internal loading of P in the STAs (i.e., the flux of soluble P from the soil to the overlying water column) can be reduced by application of soil amendments and/or soil management techniques and thereby lower outflow total P (TP) concentrations. The Detailed Study Plan (DSP) for this project specifies that work will be conducted in three phases with a STOP/GO decision made at the conclusion of the first two phases. Phase I comprised three tasks:

- Task 1 Expansion of an existing literature review on technologies for controlling soil-P flux in wetlands or lakes,
- Task 2 Data mining and/or synthesis of past South Florida Water Management District (District or SFWMD) supported projects relevant to this study and
- Task 3 To the extent practicable, assess the feasibility of implementing any of these technologies at full-scale in the STAs.

If a decision is made to continue with the study after Phase I, subsequent efforts will involve bench-scale experiments (Phase II) followed by large-scale field trials conducted in the STAs (Phase III) to document the treatment efficacy of select soil amendments and/or soil management techniques. This report summarizes the results of Phase I activities for the abovementioned tasks. Short descriptions of the Phase II and III portions of the study are provided in the Conclusions & Recommendations section and Appendix 1 of this report. Appendix I also has provisional time-lines for the start date and duration of Phase III data collection.

Task 1 - Literature Review

The Phase I literature review started with the initial set of papers and project reports compiled during the development of the Restoration Strategies Science Plan. This list was augmented with additional reference found using the web-based search engine Google Scholar¹. The focus of this effort was to generate a list of materials proven effective at sequestering P and soil management techniques that may reduce soil-P flux and potentially could be used in the STAs. This review was not intended to be a compilation of all research conducted on these materials/techniques nor to describe the biogeochemical mechanisms involved in the sequestration of P in detail. The summary provided herein relied heavily on subject matter reviews by

¹ Keywords searched in Google Scholar included "treatment wetland" AND "soil amendment", "wetland amendment", wetland AND "soil amendment", wetland AND "soil capping", wetland AND "wood chips".

CH2M Hill (2002), Douglas et al. (2004), Johansson Westholm (2006), Penn et al. (2007), Bottcher et al. (2009), Cucarella and Renman (2009), Ballantine and Tanner (2010), Ippolito et al. (2011), Vohla et al. (2011), Buda et al. (2012), and additional information found online. Reference information for journal articles, project reports and other publications was compiled into an EndNote library specific to this project (version X7, Thomas Reuters, Philadelphia, PA).

Soil Amendments

A "soil amendment" is defined for this study as any material that sorbs soluble P and includes (a) liquids, slurries or fine-grained particulate materials that can be incorporated into or broadcast on top of the soil [e.g., lime, drinking water treatment residuals (DWTRs), Phoslock[™]] and (b) course-grained materials that typically are used as the substrate in horizontal subsurface/vertical flow wetlands or in-ground/in-stream filter systems (e.g., crushed mollusk shells, various slags). Particulate materials used as soil amendments are typically rich in metal cations, primarily aluminum (Al), calcium (Ca), iron (Fe) or magnesium (Mg), that readily bond with dissolved P (Douglas et al. 2004, Johansson Westholm 2006, Penn et al. 2007, Vohla et al. 2011, Buda et al. 2012). The term "sorption" refers to adsorption and precipitation mechanisms that either separately or in combination remove dissolved constituents from solution (Penn et al. 2007, Cucarella and Renman 2009). The mode of action for Al and Fe components in particulate soil amendments is primarily through direct adsorption of P onto the particle surface whereas Ca and Mg components usually go into solution where they form P precipitates (Penn et al. 2007). The sources of soil amendments include natural materials (e.g., apatite, aragonite, gypsum, limestone, zeolite), man-made products (e.g., ferric chloride, Phoslock™, soda ash, ViroPhos[™]) and waste by-products derived primarily from manufacturing, other industrial processes or electric power generation (e.g., coal fly ash, DWTRs, Reclime[®], slags). Desirable characteristics in a soil amendment to be used in a treatment wetland include low cost, local availability in large quantities, high affinity for P, high retention of sorbed P under low redox conditions and low toxicity to flora and fauna (Penn et al. 2007, Ballantine and Tanner 2010, Vohla et al. 2011, Buda et al. 2012). Calcium and Al-based soil amendments generally are less sensitive to changes in redox conditions and thus are less likely to release bound P under the anaerobic conditions typically found in wetland soils (Ballantine and Tanner 2010).

Over 100 different materials that sorb P and have been tested for use as a soil amendment were identified during the literature review with approximately equal numbers of natural, manmade and waste by-product materials (**Table 1**). Treatment efficacy for these materials was reported as either percent P removed or sorption capacity (i.e., mass P removed/mass soil amendment). Because of marked differences in experimental methodology and study conditions that influence P sorption (e.g., water temperature and pH, amendment particle size and porosity, amendment mass to solution volume ratio, contact time of amendment with water, time required to saturate the amendment with P, initial P concentration, concentration of dissolved organic material and other anions that compete for P binding sites), it is not possible to normalize results from different studies and directly compare their sorption capacities (Johansson Westholm 2006, Cucarella and Renman 2009, Vohla et al. 2011, Buda et al. 2012). Furthermore, the majority of soil amendments have been studied only in laboratory experiments (Vohla et al. 2011, Buda et al. 2012). It is difficult to predict how these materials would perform long-term under field conditions as P retention in the laboratory can be much higher compared to rates obtained in the field (Johansson Westholm 2006, Vohla et al. 2011, Buda et al. 2012). Only long-term field trials can characterize real-world performance or reveal unfore-seen management issues for a soil amendment or soil management technique.

Based solely on the ability to sorb P, there appears to be a wide range of soil amendments that could potentially be used in the STAs (**Table 1**). However, many of these materials do not meet one or more of the other criteria desirable in a soil amendment, most notably: low cost, local availability in sufficient quantities needed for use in the STAs or no impact to downstream flora and fauna. In addition, the P removal capacity of any soil amendment is finite and once the material becomes saturated with P, the material must be replaced to restore treatment (Johansson Westholm 2006, Ballantine and Tanner 2010, Buda et al. 2012). Furthermore, the duration of most laboratory and field experiments has been too short (< 1 year) to extrapolate their results to long-term treatment efficacy. The few long-term data that are available suggest that soil amendments become saturated and lose their effectiveness within a few years after deployment and that five years may be the maximum treatment period that reasonably can be expected (Vohla et al. 2011).

Soil amendments have almost universally been tested only with domestic wastewater or agricultural runoff that had P concentrations orders of magnitude greater than STA inflow levels. Furthermore, most of the studies conducted in wetlands have been in horizontal subsurface or vertical flow systems where the soil amendment constituted most, if not all, of the substrate². Only a few studies (see CH2M Hill 2003a, Hoge et al. 2003) have attempted to use soil amendments in a fashion similar to what is proposed for the STAs, i.e., as a soil application to reduce soil-P flux thus lowering water-column TP concentrations in a free water surface wetland. Hoge et al. (2003) treated 2-ac impoundments built on organic-soil farmland with surface applications of three different soil amendments (an alum-based DWTR, gypsum and slaked lime) and monitored water-column TP concentrations in the impoundments for 16 weeks after the cells were flooded. They found that only the DWTR treatment substantially reduced water-column TP concentrations compared to the control. However, their water-column TP levels were one to two orders of magnitude greater than TP concentrations commonly observed at the outflow of the STAs. CH2M Hill (2003a) conducted a 4-month study in which they amended mesocosms

² The STAs, in contrast, are free water surface wetlands. See Kadlec and Wallace (2009) for details on the operating differences in horizontal subsurface flow, vertical flow and free water surface treatment wetlands.

filled with peat collected from STA-2 with polyaluminum chloride, ferric chloride and slaked lime. They concluded that these soil amendments were not effective in reducing the initial flux of soil P to the water column, but speculated that the short duration and small scale of the experiment may have contributed to these results.

There was discussion during the development of the Restoration Strategies Science Plan on the merits of adding wood chips in the STAs as a supplemental carbon source to enhance nutrient removal. Wood chips have been investigated in a few studies for their ability to promote nutrient removal in bioreactors and subsurface flow wetlands, but the focus of this work was on nitrogen (N) removal and not P (Christianson et al. 2009, Hopes 2010, Schipper et al. 2010, Hart 2012). The rate of denitrification in wetlands is dependent on the amount of organic carbon (C) in the soil (Reddy and DeLaune 2008). Ballantine and Tanner (2010) suggested adding wood chips in treatment wetlands where the surface layer of soil had been removed (to reduce soil-P flux) to provide a C source and promote microbial denitrification.

Soil Management Techniques

Three soil management techniques were discussed during the development of the Restoration Strategies Science Plan for reducing soil-P flux in the STAs: soil capping, soil inversion and topsoil removal. Soil capping involves placing a chemically reactive or inert material on top of the soil surface to inhibit the upward flux of P into the water column. This technique has been investigated for use in both freshwater and marine systems (e.g., Yamada et al. 1987, Jacobs and Förstner 1999, Berg et al. 2004, Ye et al. 2006, Chimney et al. 2007, Hickey and Gibbs 2009, Lin et al. 2011, Ding et al. 2012, Meis 2012, Meis et al. 2012, Xu et al. 2012). One spoil-capping approach, alum application, has been widely used to inactivate sediment P in the hypolimnion of shallow eutrophic lakes (Welch et al. 1988, Welch and Cooke 1999, Cooke et al. 2005). As noted in the previous section, Hoge et al. (2003) applied soil amendments to the soil surface in test impoundments to inhibit soil-P flux. Soil inversion, which involves plowing or disking the nutrient-rich topsoil layer so that it is replaced with the underlying nutrient-poor subsurface soil horizon, has been proposed as a way to manage nutrient export in agricultural systems (Daniel et al. 1998, Pekrun et al. 2003). Soil removal is a practice used in lake restoration to reduce internal nutrient loading (Cooke et al. 2005). Removal of the nutrient-rich surface soil layer, thereby exposing the lower-nutrient subsoil (or limestone caprock), has been suggested as a method to eliminate a source of internal P loading in treatment wetlands (Ballantine and Tanner 2010, Lindstrom and White 2011).

Task 2 – Relevant District-supported Projects

The District has carried out a number of projects that are relevant to this study. Chimney et al. (2007) performed short-term laboratory experiments that found broadcasting a soil amendment (Reclime[®]) on top of sediment cores (peat collected from the footprint of STA-3/4) was much more effective at reducing soil-P flux to the overlying water column immediately after the cores were flooded versus the inhibition of soil-P flux achieved with mixing Reclime® into the soil. The District has evaluated the effectiveness of soil capping. In the largest platform tested to date, a 5-ac wetland cell capped with limerock at the Field-scale PSTA Research Site achieved a flow-weighted mean (FMW) outflow TP concentration of 18 µg L⁻¹ during a 9month period of optimal performance (CH2M Hill 2003b). The District has investigated the potential benefits of soil removal to reduce the internal P loading (Reddy et al. 2002) and has tested the effectiveness of this approach in the field with two different platforms. In the first study, another 5-ac cell at the Field-scale PSTA Research Site was scraped down to the limestone caprock and achieved a FWM outflow TP concentration of 16 µg L⁻¹ during the same operational period noted above (CH2M Hill 2003b). In the second study, the 100-ac PSTA cell in the STA-3/4 PSTA Project also was scraped down to the limestone caprock and had annual FWM outflow TP concentrations that ranged from 8 to 12 µg L⁻¹ for six consecutive operational years (Zamorano et al. 2014). The District has conducted a number of pilot projects that demonstrated the utility of inverting the topsoil with the subsurface soil horizon to reduce copper concentrations at the soil surface (Environmental Consulting & Technology 2003, 2006, Shaw Environmental 2007, Water and Soil Solutions 2009, WRScompass 2009, South Florida Water Management District and URS 2014) or to reduce the flux of soil-P to the water column (South Florida Water Management District 2009).

Task 3 - Feasibility Assessment

The feasibility of using soil amendments or soil management techniques to reduce outflow TP concentrations in the STAs can be evaluated from several different perspectives: constructability, treatment efficacy, operations/regulatory issues and economics. Not all these areas can be fully addressed in Phase I of this study.

Constructability

The District has experience with the large-scale engineering and construction practices that would be required to employ soil amendments or soil management techniques in the STAs. For example, during the rehabilitation of STA-1W and STA-5, a number of treatment cells were dewatered while the remainder of the STA remained in operation, the vegetation and accrued sediment in the dewatered cells were removed, or the cell topography was recontoured, and the aquatic plant communities reestablished after the cells were reflooded. The District also

has experience with soil capping³, topsoil removal⁴ and soil inversion projects⁵ at a large scale. Therefore, there is no reason from a constructability perspective that soil amendments or soil management techniques could not be employed at full-scale in the STAs.

Treatment Efficacy

Despite the number of studies focused on soil amendments and soil management techniques (see **Table 1**), no published data were found that demonstrated the long-term efficacy of using these approaches to reduce outflow P concentrations in free water surface treatment wetlands. While soil amendments and soil management techniques, in theory, may enhance STA treatment performance, field trials will be necessary to verify their long-term usefulness. Therefore, a meaningful assessment of the treatment efficacy of any of these approaches in the STAs cannot be made at this time.

Operations & Regulatory Issues

There are a number of potential operations and regulatory issues related to using soil amendments or soil management techniques in the STAs that need to be addressed. These issues and associated questions are presented in **Table 2**. The ramifications of using soil amendments and/or soil management techniques in the STAs must be considered to help guide the STOP/GO decision for this study after the completion of Phase I, and determine if the study team should continue with further investigation and decide whether there will be any constraints on conducting this research.

Economics

Cost were estimated for: (a) the construction and infrastructure associated with large-scale test facilities (LSTFs) needed to conduct field trials in the STAs and (b) implementing soil amendments and soil management techniques at full-scale in the STAs. The original conceptual design for the LSTFs has been revised and now includes an option that utilizes two cells in the STA-1W Expansion Area (EA) to test the efficacy of soil inversion (Field-trial Option #1) and a second option that calls for building four sub-cells in each of four STAs (STA-1E, STA-1W, STA-3/4 and STA-5/6) to test several different technologies (Field-trial Option #2; see **Appendix 1** for details).

³ District field-station personnel built two 5-ac limerock pads (12 and 24 inches thick) within Cell 3 of STA-2 and a 5-ac limerock-capped cell (24 inches thick) at the Field-scale PSTA Research Site adjacent to STA-2. This later site was decommissioned and is now part of STA-2 Cell 4.

⁴ The District, or its contractors, removed all soil down to the caprock in another 5-ac cell located at the Field-scale PSTA Research Site and in the 100-ac PSTA Cell located at the STA-3/4 PSTA Project.

⁵ Soil inversion pilot projects were conducted by Environmental Consulting & Technology (2003, 2006), Shaw Environmental (2007), South Florida Water Management District (2009), Water and Soil Solutions (2009), WRScompass (2009), South Florida Water Management District and URS (2014).

Costs for Field-trial Option #1 would only entail installing autosamplers at the inflow and outflow water control structure of Cells 7 and 8 should this equipment not be part of the STA-1W EA design (\$178K, **Table 3**). In Field-trial Option #2, the application of one soil amendment, limerock capping and soil inversion in the sub-cells in each abovementioned STA⁶ will be tested. The cost of the soil amendment was derived from the range of prices for hydrated lime (calcium hydroxide [Ca(OH)₂]) found online and a field application rate of 3 t/ac (**Table 4**). The other costs for this option were developed with the assistance of Mr. Jack Ismalon, Principle Cost Estimator, Engineering and Construction Bureau, SFWMD.

Estimated construction and infrastructure costs for the LSTFs in Field-trial Option #2 ranged from a low of \$3.7M in STA-1W to a high of \$24.8M in STA-5/6; the total estimated construction cost for all four STAs is \$57.5M (Table 4). The estimated costs to apply a soil amendment, install a 6-inch thick limerock cap or invert the surface layer of soil to a depth of 2 ft over the entire STA surface area dominated by SAV (31,284 ac) are \$99.1M, \$876.4M and \$85.6M, respectively (Table 5). As noted above, soil amendments eventually become saturated with P and will need to be replaced on a periodic basis⁷. A limerock cap may require periodic maintenance at some indeterminate frequency to remove the new sediment, which may be P rich, that accrues on the limerock cap surface. Reapplication of a soil amendment or maintaining a limerock cap will be additional costs that recur throughout the operational life of an STA; note that these additional costs are not captured in **Table 5**. The cost estimates in **Tables 4** and **5** are preliminary and there is uncertainty in the assumptions used to generate them. As such, all costs should be regarded as order of magnitude estimates and that a wide range of values could be generated by varying one or more of the scenario assumptions. The cost threshold at which any soil amendment or soil management technique becomes economically infeasible for the District to implement in the STAs is outside the scope of this report.

Conclusions & Recommendations

Our conclusions and recommendations for this Phase I portion of the study are as follows:

 While all of the soil amendments listed in **Table 1** sorb P to varying degrees, many of them are not suitable for use in the STAs primarily because they are either not available locally, or if available, could not be supplied in sufficient quantities to meet our needs. A secondary consideration is potential negative impacts to downstream Everglades flora and fauna, which are unknown for most of these materials. For Fieldtrial Option #2 (Appendix 1), our recommendation is to test one or more agricultural

⁶ One sub-cell in each STA will be operated as a control and will not be manipulated. We elected not to test adding wood chips because there is no evidence in the literature that this technique is effective at reducing P. We elected not to test topsoil removal because any significant lowering of the bottom elevation in the STAs likely would adversely affect system hydraulics.

⁷ The frequency of soil amendment refurbishment is unknown at this time; our best estimate is that it will be required every 3 to 5 years.

products that are known to sorb P (e.g., aragonite or Reclime[®]), are available in sufficient quantities for use in the STAs and have low potential for negative impacts to downstream flora or fauna.

- 2. All of the operations and regulatory issues raised in **Table 2** should be considered before a decision to proceed with the next phase of this study is made. This study should continue beyond Phase I only if the District is reasonably certain that the use of soil amendments or soil management techniques would be acceptable in the STAs and the receiving Everglades water bodies.
- 3. The DSP for this project specifies that it will be conducted in three sequentially ordered phases: Phase I \rightarrow Phase II \rightarrow Phase III. Phase II was to consist of small-scale experiments that screen a variety of soil amendments and/or soil management techniques identified in Phase I for their ability to inhibit soil-P flux. Phase III field trials would then test the most promising of these soil amendments/management techniques at large scale. However, based on what was learned in Phase I about the P sorption capabilities of various materials and the limitations of transferring laboratory results to the field, coupled with the SFWMD's recent experience testing soil amendments (Chimney et al. 2007, Chimney et al. 2013), there is no longer a need to conduct a large number of experiments with multiple soil amendments. Instead, if a GO decision to continue with the study after Phase I is reached, a modified Phase II to identify the optimum application rate(s) for a few select soil amendments (e.g., aragonite and Reclime®) will be conducted. Concurrent with the start of Phase II work, Phase III will be initiated and design of the LSTFs that are described in Appendix I will begin. Limerock capping and soil inversion also will be tested in these LSTFs. Given the substantial cost of implementing a soil amendment or soil management technique at full-scale in the STAs, and the inherent biogeochemical variability among STAs (see Table 5), it is recommended that the field trials be conducted in the four STAs specified in Appendix 1 to demonstrate that these technologies will work across all the STAs.

STOP/GO Decision

This Phase I Summary Report was reviewed by the Restoration Strategies (RS) Science Plan Management Team (Team). Considering the uncertainties in treatment efficacy, potential impacts to STA operations and the economics associated with conducting large-scale field trails and implementing any of these technologies at full-scale in the STAs, the Team recommended that (a) the study move forward with planning associated with Field-trial Option #1 in the STA-1W EA and (b) not to proceed with study Phases II and III for Field-trial Option #2 at this time. Cells in the STA-1W EA are scheduled to be flow-capable on December 31, 2018. The RS Steering Committee on October 6, 2015 agreed with these STOP/GO recommendations.

Table 1. Soil amendments and soil management techniques identified in a literature review con-
ducted for Phase I of the Use of Soil Amendments/Management Techniques to Control
P Flux Study.

| Soil Amendment/ | | | |
|------------------------------|------------|---|-------------------------|
| Management Technique | Туре | Description/Composition/Characteristics | References [§] |
| acid mine drainage resid- | by-product | A waste product generated during the neutralization of | 47 |
| uals | | acid mine drainage | |
| AlgalBLOCK | man-made | A precipitated form of surface activated calcium car- | 19 |
| | | bonate | |
| allophane | natural | An aluminum silicate clay mineraloid [Al ₂ O ₃ ·(SiO ₂) _{1.3-} | 5, 20 |
| | | $_2$ ·(2.5-3)H ₂ O] that contains alumina; found in volcanic | |
| | | and non-volcanic derived soils | |
| alum/aluminum sulfate | man-made | $AI_2(SO_4)_3 \cdot 14 H_20$; alkaline - highly caustic & reactive; low | 2, 4, 5, 6, 9,14, 17, |
| | | solubility | 25, 36, 37, 39, 40, |
| - Luncher | | | 47, 48, 62 |
| alumina alumina-activated | natural | Aluminum oxide [Al ₂ O ₃] | 35 20 |
| alumina-activated | man-made | Aluminum oxide [Al ₂ O ₃] A waste material produced during aluminum ore pro- | 35 |
| | by-product | cessing; contains a mixture of Ca, Al, Fe and Al-coke | 55 |
| aluminum chloride | man-made | AICl ₃ | 6, 47 |
| alunite | man-made | Hydrated aluminum potassium, sulfate mineral [KAl ₃ (SO ₄) ₂ (OH) ₆] | 58 |
| apatite/phosphate rock | natural | A calcium phosphate mineral [Ca10(PO4)6(OH, F, Cl)2] with high concentrations of Ca, OH, F and Cl | 5, 58 |
| aragonite | natural | A calcium carbonate mineral precipitated from sea water; | 12 |
| | | evaluated during the New Alternative Technology As- | |
| | | sessment (NATA) Program (see Chimney et al. 2013) | |
| Baraclear® | man-made | A proprietary mixture of alum and other non-toxic earth- | 5 |
| | | en materials; manufactured as nodules with a diameter of | |
| | | 1/4" to 3/8" | |
| bauxite | natural | An aluminum ore rich in hydrated Al and Fe oxides | 16, 21, 32, 58 |
| bauxite mine waste | by-product | A waste material produced during the processing of baux- ite ore | 47 |
| Bauxsol™ | by-product | Neutralized bauxite residuals | 58 |
| bentonite | natural | An absorbent aluminum phyllosilicate clay; contains Ca, Mg, Fe, Al and Si oxides | 16 |
| black oxide | by-product | A waste material derived from mineral sands processing | 16 |
| bone char | man-made | A granular material produced by charring animal bones | 44 |
| brick (Fe-coated) | man-made | | 16 |
| calcareous rock (crushed) | natural | | 3 |
| calcium carbonate/calcite | by-product | A carbonate mineral; the most stable form [CaCO ₃]; re- | 2, 9, 17, 20, 36, |
| | | sidual precipitated during drinking water treatment | 48, 62 |
| calcium chloride | man-made | CaCl ₂ ; produced from limestone or during the manufac- ture of soda ash | 37, 47 |
| calcium silicate slag | by-product | | 47 |
| cement kiln dust | by-product | A fine dust-sized material resulting from cement produc- tion; composed primarily of CaO and SiO ₂ | 44, 47 |
| coal fly ash | by-product | A waste product from burning coal; contains silica diox- | 5, 16, 20, 21, 32, |
| | | ide, alumina and ferric oxides | 46, 47, 58 |
| coal fly ash (anthracite) | by-product | A waste product from burning anthracite coal; contains amorphous ferric hydroxide [Fe(OH) ₂] | 8 |
| | | | |

| Soil Amendment/ | | | |
|--------------------------------|------------|--|---|
| Management | | | |
| Technique | Туре | Description/Composition/Characteristics | References§ |
| coal fly ash (bitumi- nous) | by-product | A waste product from burning bituminous coal; contains calcium hydroxide [Ca(OH) ₂] and calcium sulfate [CaSO ₄ ·2H ₂ O] | 8 |
| copperas | man-made | a.k.a. ferrous sulfate or iron(II) sulfate [FeSO4·7H2O]; available in granular form | 6 |
| Damolin (new product) | man-made | A processed calcium carbonate product | 3 |
| Damolin (old product) | man-made | A processed calcium carbonate product | 3 |
| dinoSoil | natural | A leonardite containing humic and fulvic acids, montmo- rillonite clay, Fe & Al oxides and other minerals; from Texas | 46, 57 |
| dolomite/dolomite sand | natural | Calcium magnesium carbonate [CaMg(CO ₃) ₂]; see lime (agricultural/limerock) | 2, 32, 58 |
| DWTR | by-product | Drinking water treatment residuals derived from alumi- num, calcium or iron salts; also may have some polymers | 5, 9, 17, 47, 62 |
| DWTR-AI | by-product | The residuals of Al salts used in the treatment of drinking water | 1, 6, 8, 24, 29, 31, 38, 39, 44, 45, 46, 48 |
| DWTR-Ca | by-product | The residuals of Ca salts used in the treatment of drinking water | 24, 35, 46 |
| DWTR-Fe | | | 6, 24, 31, 35, 38, 46 |
| ferric chloride | man-made | Iron(III) chloride [FeCl ₃ ·6H ₂ O]; strong acid | 2, 6, 9, 10, 17, 33, 37, 47, 62 |
| ferric oxide | natural | | 20 |
| ferric sulfate | man-made | Iron(III) sulfate [Fe ₂ (SO ₄) ₃ ·3H ₂ O]; strong acid | 6, 9, 47 |
| Filtra P | man-made | Produced by heating limestone, gypsum and Fe oxides; high calcium hydroxide content | 58 |
| Filtralite-P™ | man-made | An amended clay material developed for P removal by sorption to Al, Ca, Mg and Fe oxides | 3, 5, 16, 20, 32, 58 |
| GAC | man-made | Granular activated charcoal | 44 |
| gravel | natural | Trace amounts of Ca, Mg, Fe and Al oxides | 16, 20, 58 |
| gypsum | natural | Hydrated calcium sulfate[CaSO ₄ ·(2H ₂ O)]; neutral pH | 8, 9, 15, 16, 29, 46, 47, 55 |
| gypsum (waste) | by-product | A waste product produced during the processing of gyp- sum | 47 |
| HeloFIR® | man-made | A dark granular material specific for P removal | 3 |
| HiClay [®] alumina | by-product | A proprietary material generated during alum production & other bauxite-based processes | 9 |
| humate product | by-product | A dried waste material produce during titanium mining | 35 |
| hydrotalcite | natural | A layered double hydroxide soil mineral of the general formula $(Mg_6Al_2(CO_3)(OH)_{16}\cdot 4(H_2O)$ | 20 |
| Hyper+ion 1090 | man-made | A polyaluminum hydroxychloride solution | 6 |
| imogolite | natural | An aluminum silicate clay mineral [Al ₂ SiO ₃ (OH) ₄]; occurs in soils formed from volcanic ash | 20 |
| iron ore | natural | | 58 |
| iron oxides/steel wool | man-made | Materials added to peat and sand to enhance P removal | 5 |
| lanthanum | natural | A rare earth element | 20 |
| laterite | natural | A soil type rich in Fe and Al; formed in hot and wet tropi- cal areas; derived from decomposition of bauxite rock | 32, 58 |

| Technique | Туре | Description/Composition/Characteristics | References [§] |
|----------------------------------|------------|--|---|
| LECA® | man-made | Light-Expanded Clay Aggregates; an amended clay mate- rial developed for removing P by sorption to Al, Ca, Mg and Fe oxides | 5, 16, 20, 21, 32, 58 |
| Lehigh cement | man-made | Contains Ca, Mg, Fe and Al | 16 |
| lime | | | 6, 46 |
| lime (agricultural/ limerock) | natural | a.k.a. dolomite [CaMg(CO ₃) ₂]; weakly alkaline | 9 |
| lime (mixed) | by-product | A mixture of CaO, Ca(OH) ₂ and CaCO ₃ ; spent material generated during sugar refining | 34 |
| lime (quick/burnt) | man-made | Calcium oxide [CaO]; strongly alkaline and reactive | 9, 34, 47 |
| lime (slaked/hydrated) | man-made | Calcium hydroxide [Ca(OH) ₂]; alkaline with low solubility | 2, 4, 9, 10, 29, 47, 48 |
| lime (sugarbeet) | by-product | A waste material produced from sugarbeets during sugar production | 62 |
| limestone/limerock | natural | A sedimentary rock composed largely (> 50%) of calcium carbonate [CaCO3] | 5, 16, 17, 18, 21, 28, 32, 41, 44, 47, 56, 58 |
| LWA | man-made | Light-weight aggregates; an amended clay material de- veloped for P removal by sorption to Al, Ca, Mg and Fe oxides | 58 |
| maerl | natural | Collective name for deposits of calcareous red algae; dredged from the sea floor; high CaCO ₃ content | 26, 32, 58 |
| mag dust | by-product | A waste product derived from building practices | 55 |
| marble | natural | A non-foliated metamorphic rock composed of recrystal- lized carbonate minerals, most commonly calcite (lime- stone) or dolomite | 58 |
| marl gravel | natural | A solidified soil consisting of clay and fine particles of limestone | 32, 58 |
| Nclear [®] | man-made | A proprietary mixture of calcium silicate hydroxides; evaluated during the NATA Program | 12 |
| Norlite | man-made | A lightweight course aggregate made from fired shale | 28, 58 |
| ochre | by-product | A waste product from treatment of abandoned mine waters containing hydrated iron oxide [Fe(OH)3 & FeO(OH)] | 5, 58 |
| ochre (black) | by-product | A waste product from treatment of abandoned mine waters; has a high Al content | 55 |
| ochre (red) | by-product | A waste product from treatment of abandoned mine waters; has a high Al content | 55 |
| oil shale (burnt) | by-product | A waste product from heating oil shale to produce miner- al oil | 21, 32, 58 |
| oil shale ash | by-product | Residue remaining after the combustion of oil shale in Estonian thermal power plants | 58 |
| opoka | natural | An amorphous calcareous sedimentary rock found in south-eastern Poland; contains Ca, Mg, Fe, Al and Si ox-ides | 16, 32, 58 |
| PACL | man-made | Polyaluminum chloride (polyaluminum hydroxychloride) [Al ₂ (OH) _n Cl _{6-n} ·nH ₂ O)]; mildly acidic | 6, 9, 10, 33, 39 |
| PAM | man-made | Polyacrylamide (PAM) polymer: anionic or cationic; neu- tral pH | 6 |

| Management | | | |
|---------------------------------|------------|---|---------------------------------|
| Technique | Туре | Description/Composition/Characteristics | References§ |
| paper mill sludge | by-product | A by-product of paper production; contains Al and Ca | 47 |
| peat | natural | A soil that has high organic matter content; accumulation of partially decayed vegetation or organic matter | 18, 58 |
| | | 5, 12, 19, 42, 43 | |
| PNAS | man-made | Partially neutralized aluminum sulfate | 39 |
| polonite | man-made | Manufactured by heating opoka to 900 °C; consists main- ly of reactive lime and wollastonite phases | 58 |
| polymers | man-made | This category includes a number of different anionic, cationic and non-ionic products | 9, 25 |
| Pro-Sil | by-product | A silica slag containing Ca, Ma, Si and other micronutri- ents | 46 |
| pumice | natural | A porous, low-density rock produced when volcanic lava with a high water and gas content cools and hardens | 5 |
| pumice soil | natural | A light, porous soil of volcanic origin with high levels of Al, Fe, Ca and Mg; can have varying amounts of allophane | 5 |
| Reclime® | by-product | A Ca-Mg silicate with impurities; by-product of steel pro- duction; alkaline | 9, 11 |
| red mud | by-product | A waste material generated during alumina refining from bauxite | 20, 58 |
| sand | natural | This category encompasses a wide variety of sand types with differing physico-chemical properties | 5, 16, 20, 58 |
| sand (Al/Fe-coated) | man-made | | 16, 35 |
| sand (concrete) | natural | | 35 |
| sand (foundry) | by-product | A waste material from sand molds used in metal casting | 47 |
| sand (masonry) | natural | | 35 |
| sand (organic) | natural | | 35 |
| sand (quartz) | natural | | 18, 58 |
| sandblast grit | by-product | A waste product from sandblasting | 35 |
| sander dust | by-product | A waste product from building practices | 55 |
| serpentinite | natural | A magnesium-rich silicate mineral associated with igne- ous rocks [H ₄ Mg ₃ Si ₂ O ₉] | 5 |
| shale | natural | An argillaceous, fine-grained sedimentary rock derived from limestone | 5, 16, 21, 58 |
| shells | | | |
| shells (crushed blue mussel) | by-product | Waste shells produced by aquaculture; high Ca content | 3 |
| shells (crushed sea- shells) | by-product | | 3, 5, 32 |
| shells (oyster) | by-product | Waste shells produced in oyster culture; high Ca content | 58 |
| shellsand | natural | A natural calcareous material formed mainly from crushed shells, snails and coral algae | 5, 32, 58 |
| siderite | natural | A mineral composed of iron(II) [ferrous] carbonate [FeCO ₃] | 8 |
| slag | by-product | By-product of steel and iron production containing vary- ing amounts of Ca, Mg, Al, Si and Fe oxides | 5, 8, 16, 20, 32, 41, 44, 58 |
| soda ash | man-made | Sodium carbonate [Na ₂ CO ₃] | 6 |
| sodium aluminate | man-made | Na ₂ Al ₂ O ₄ ; weakly alkaline and extremely reactive | 6, 9, 14, 50 |
| sodium bicarbonate | man-made | NaHCO ₃ | 6 |

| Soil Amendment/ Management | | | |
|-------------------------------|------------|--|---------------------|
| Technique | Туре | Description/Composition/Characteristics | References§ |
| sodium hydroxide | man-made | a.k.a. caustic soda or lye [NaOH] | 6 |
| sodium nitrate | natural | a.k.a. saltpeter; NaNO ₃ | 37 |
| soil | natural | P retention is related to the mineral content of Fe and Al | 20, 58 |
| 5011 | naturai | compounds | 20, 30 |
| soil (calcareous) | natural | A soil with a high $CaCO_3$ content | 16 |
| soil (glossic hapludalf) | natural | A fine loamy, mixed soil formed from calcareous and | 28 |
| 66 (8.666.6ap.a.da.) | | sulfur-rich glacial till | |
| soil (spodosol) | natural | Ashy gray, acidic soils with a strongly leached surface | 16, 32, 58 |
| | | layer; the B horizon of a forest soil; contains Fe and Al | |
| | | oxides | |
| soil capping | - | Cover existing soil with a layer of material with low P | 19, 42, 43, 61 |
| | | content to reduce P flux from the soil to the water col- | |
| | | umn | |
| soil inversion | - | Disc/plow so as to place the high nutrient/contaminate | 22, 23, 52, 53, 54, |
| | | topsoil layer underneath the deeper sub-soil layer | 59, 60 |
| soil removal | - | Remove topsoil containing the highest nutrient/con- | 5, 36, 49 |
| | | taminate concentrations; may restrict N removal due to | |
| | | initial lack of organic matter | |
| STI | man-made | Simtec triad ionate; a proprietary clay-like mineral prod- | 12 |
| | | uct that contains Al, Fe, Ca and Mg; evaluated during the | |
| | | NATA Program | |
| SuperMag | by-product | A magnesium-based fertilizer by product | 35 |
| Tennessee slag | by-product | A Ca-Mg silicate with impurities; by-product from electric | 9, 11 |
| | | furnace production of P; alkaline | |
| tephra | natural | The fragmented material explosively erupted from a vol- | 5 |
| | | cano; may contain large amounts of allophane | |
| titanium mine waste | by-product | A waste material produced during the processing of tita- | 46 |
| | | nium ore | |
| utelite | man-made | A lightweight expanded shale aggregate | 16 |
| vermiculite | natural | A hydrous, silicate mineral classified as a phyllosilicate; | 55 |
| | | (Mg ⁺² , Fe ⁺² , Fe ⁺³) ₃ [(AlSi) ₄ O ₁₀] ·(OH) ₂ ·4H ₂ O] | |
| ViroPhos™ | man-made | Proprietary mixture of hematite (Fe2O ₃), alumina | 12 |
| | | $(AI(OH)_3)$, sodalite $(Na_4(AI_6Si_6O_{24})CI_2)$ and quartz (SiO_2) ; | |
| | | evaluated during the NATA Program | |
| wollastonite | natural | A calcium inosilicate (metasilicate) mineral composed of | 5, 7, 8, 9, 18, 28, |
| | | calcium oxide and silicon dioxide [CaSiO ₃] | 32, 58 |
| wood (ash) | by-product | | 47 |
| wood (chips) | natural | Investigated for N removal or neutralization of acid mine drainage | 13, 27, 30, 51 |
| wood (treebark) | natural | | 5 |
| woodchip biochar | man-made | Investigated for N removal | 27 |
| WP-1™ | man-made | A proprietary mixture of mineral compounds; evaluated during the NATA Program | 12 |
| zeolite | natural | Microporous aluminosilicate minerals with low bulk den- sity and high porosity; can contain Ca and Mg | 5, 16, 21, 32, 58 |
| zoolitor (amondod) | man made | | 5 |
| zeolites (amended) | man-made | Zeolite pre-treated with cationic surfactants to enhance anion retention | 5 |

[§]Reference key: [1] Agyin-Birikorang et al. (2009), [2] Ann et al. (2000), [3] Arias & Brix (2004), [4] Babin et al. (1992), [5] Ballantine & Tanner (2010), [6] Bottcher et al. (2009), [7] Brooks et al. (2000), [8] Callahan et al. (2002), [9] CH2M Hill (2002), [10] CH2M Hill (2003a), [11] Chimney et al. (2007), [12] Chimney et al. (2013), [13] Christianson et al. (2009), [14] Cooke et al. (1981), [15] Cox et al. (2005), [16] Cucarella & Renman (2009), [17] DB Environmental, Inc. (2013), [18] DeBusk et al. (1997), [19] Ding et al. (2012), [20] Douglas et al. (2004), [21] Drizo et al. (1999), [22] ECT, Inc. (2003), [23] ECT, Inc. (2006), [24] Elliott et al. (2002), [25] Florida Department of Environmental Protection (2011), [26] Gray et al. (2000), [27] Hart (2012), [28] Hill et al. (2000), [29] Hoge et al. (2003), [30] Hopes (2010), [31] Ippolito et al. (2011), [32] Johansson Westholm (2006), [33] Jorge et al. (2002), [34] Kirkkala et al. (2012), [35] Leader et al. (2008), [36] Lindstrom & White (2011), [37] Liu et al. (2009), [38] Makris et al. (2005), [39] Malecki-Brown & White (2009), [41] Mara et al. (2007), [42] Meis (2012), [43] Meis et al. (2012), [44] Mortula et al. (2007), [45] Novak & Watts (2005), [46] O'Connor et al. (2005), [47] Penn et al. (2007), [48] Reddy et al. (1998), [49] Reddy et al. (2002), [50] Sanville et al. (1976), [51] Schipper et al. (2010), [52] South Florida Water Management District & URS (2014), [53] SFWMD (2009), [54] Shaw (2007), [55] Spears et al. (2013), [56] Strang & Wareham (2006), [57] Struve & Zhou (undated), [58] Vohla et al. (2011), [59] Water and Soil Solutions, LLC (2009), [60] WRScompass, Inc. (2009), [61] Xu et al. (2012), [62] Zvomuya et al. (2006).

Table 2. Feasibility issues, questions and comments generated during Phase I of the Use of Soil Amendments/Management Techniques to Control P Flux Study that require consideration before the study moves to the next phase.

| | Feasibility Issue/Question/Comment Response/Resolution* | | | | | | |
|---|---|---|--|--|--|--|--|
| 1 | Feasibility Issue/Question/Comment Limerock capping involves adding a layer of crushed limerock (i.e., limestone) on top of the existing soil in the downstream treatment cells of the STAs. The depth of the limerock layer that will be required to provide treatment is unknown at this time, but may range from several inches to more than one foot. Adding limerock will reduce a cell's storage capacity and may incur a hydraulic penalty, i.e., the increase in surface elevation will impede flow as water coming from upstream treatment cells would have to move "uphill" as it enters capped cells. Two operational issues that may constrain the thickness of the limerock layer need to be addressed: a. How much loss of storage capacity in a cell due to capping can be tolerated? b. How much of a hydraulic impediment to flow in a cell due to capping can be tolerated? | Response/Resolution* While the loss of cell storage ca- pacity and flow restriction associ- ated with limerock capping were recognized as problems in the STAs, an upper tolerance limit for either issue that would preclude the use of this technology was not identified. However, due to the high projected cost of installing a limerock cap, this approach is the least attractive of the treatment technologies included in this study. | | | | | |
| 2 | Applying a soil amendment or installing a limerock cap in the STAs will require that treatment cells be dewatered and the vegetation removed before the work is done. These cells would then be reflooded and the wetland plant community allowed to reestablish itself before flow-through operations could resume. This process will take at least one year (and possibly longer) during which time the entire flow-way would be offline. Any soil amendment eventually will become saturated with P and have to be replaced. New soil will accrue on top of a limerock cap over time; if this soil is P rich, it may have to be removed to prevent P flux back to the water column. Removing this soil will likely disturb the upper portion of the limerock cap and necessitate that it be repaired. Reapplication of a soil amendment or maintenance of a limerock cap will be required throughout the operational life of a STA at some yet undetermined frequency and require that cells again be taken offline for a least on year while the work is done and the plant community becomes reestablished. Consider that every STA flow-way may need treatment to achieve the WQBEL criteria. Some issues that need to be addressed concerning using soil amendments or a limerock cap in the STAs include: a. Can STA flow-ways be taken off-line for one year (or longer) for the initial application of a soil amendment or installation of a limerock cap? b. Based on the operating scenario assumed for the STAs during the design of the Restoration Strategies projects, how frequently can a flow-way be taken offline in subsequent years to reapply the soil amendment or perform maintenance on the limerock cap? | Restoration Strategies assumes that each STA will be taken offline entirely for maintenance activities once every 20 years. While taking STA flow-ways offline more fre- quently would not be ideal, doing so for the initial installation of a soil amendment or limerock cap and subsequent reapplication of the soil amendment or mainte- nance activities on the limerock cap at a shorter time interval in future years could be tolerated. | | | | | |
| 3 | Applying a soil amendment or limerock cap has been criticized as not economically feasible to implement at full-scale in the STAs due to their size. However, the dollar threshold above which these technolo- gies become cost prohibitive is undefined. The affordability of any of these technologies ultimately is a policy issue. | This is just a comment; no input was elicited from the focus group. | | | | | |

Table 2. (continued).

| | Feasibility Issue/Question/Comment | Response/Resolution |
|---|--|--|
| 4 | Application of a soil amendment in the STAs more than likely will re- quire enormous quantities of material, i.e., many tens of thousands of tons per application. A consideration in selecting any particular prod- uct for testing is the availability of sufficient quantity needed for use over the operational life of the STAs. | Our inclination at this time is to test an agricultural product that is 1) widely used in the EAA, 2) available locally in quantities that will meet our future needs for use in the STAs and 3) has no toxicity issues for downstream Everglades flora and fauna |
| 5 | The application of a limerock cap or a soil amendment in the STAs may have regulatory implications: a. Would using either of these approaches in the STAs require a change to the current STA operating permit? b. If so, what would USEPA and FDEP require from the District in order to authorize their use in the in the STAs? Will the District have to demonstrate that outflow from a cell treated in either fashion is "marsh ready"? c. Will the District be allowed to divert water around the STAs when flow-ways are off-line for soil amendment reapplication or maintenance of the limerock cap? | All regulatory issues will need to be addressed with USEPA and FDEP prior to implementing any of the technologies that are part of this study. However, none of the issues raised here is seen as prohibiting the use of limerock capping or soil amendment in the STAS. |
| 6 | Phase III of this study calls for conducting field trials of candidate technologies in the STAs. Our initial proposal to construct large enclosures (~ 10 ac) within the STAs for this work was criticized by the Technical Representatives as being too small and subject to experimental artifacts. There is agreement (in part) with their assessment. Alternate suggestions have been to use entire treatment cells for testing. One approach would involve constructing new longitudinal internal levees to partition existing cells and conduct tests within these new sub-cells. Issues that need to be addressed concerning this approach include: a. Is such an approach feasible from an operations or regulatory perspective, i.e., can the entire treatment cells be used for testing? b. There is no guarantee that any soil amendment/management technique shown to work in one STA will necessarily work when applied to the other STAs. Ideally, tests would be conducted in every STA. Is there enough operational flexibility in the STAs? c. Conducting experiments in entire cells has risks. There is no guarantee that treatment performance in cells after they are manipulated will improve relative to pre-testing performance and there is a possibility that post-manipulation performance may be worse than pre-testing performance, i.e., one or more cells could "break" during testing. There also may be other unanticipated negative impacts. In addition, returning manipulated cells to their original condition after the study may require restarting the cell (i.e., removing the new levees and the amended soil or limerock cap and reestablishing vegetation) or will not be possible in the case of soil inversion. | It is recognized that there is po- tential risk in conducting research on these technologies at a large- scale in the STAs. Nevertheless, these concerns would not pre- clude conducting large-scale field trials to test these technologies. |

Table 2. (continued).

| | Feasibility Issue/Question/Comment | Response/Resolution |
|---|---|--|
| 7 | How will success be judged in this study, i.e., what level of treatment performance will the field trials have to achieve to be judged as having been successful? Will it be sufficient for treated cells to produce an outflow TP concentration that is only measurably lower than the con- trol cell or will the treated cells have to achieve outflow TP levels that meet the WQBEL criteria? | Implementing one of these tech- nologies ideally would enable the STAs to achieve the WQBEL. However, enhanced treatment performance that falls short of the WQBEL may be acceptable under some circumstances. |
| 8 | If a soil amendment or a soil management technique were to be implemented in the STA-1W Expansion Area or in the A-1 Flow Equalization Basin (FEB), it would be easier to install the technology during construction when the work can be done in the dry. However, based on the anticipated construction schedule for the STA-1W Expansion Area and the A-1 FEB, there may not be sufficient time to finish the field-trial phase of this study, and validate that the technology works, before installation of the technology would have to start. Would the District want to commit to implementing a technology in the STA-1W Expansion Area or the A-1 FEB before the final phase of this study has been completed? | Implementing any of the technol- ogies that are part of this study throughout the entire STA-1W Expansion Area is not being con- sidered at this time. |
| 9 | One soil management technique that has been proposed for treatment wetlands is topsoil removal. The idea is to remove a source of P that can flux back to the water column. However, topsoil removal would lower the wetland floor and may affect system operation. If topsoil is removed from the STAs, how much material can be taken out before system hydrology and hydraulics are compromised, i.e., how far below the original soil elevation can excavation occur before there would be problems moving water through and maintaining a target stage in the STAs? | Due to hydraulic issues associated with lowering the ground eleva- tion in the STAs, topsoil removal has been excluded from the tech- nologies that will be studied as part of this study. |

* Input to these issues and questions was provided by a focus group that consisted of Lawrence Gerry, Susan Gray, Delia Ivanoff, Jill King, Jennifer Leeds, Jeremy McBryan, Kim O'Dell and Larry Schwartz.

| | C | ell 7 | Ce | Cell 8 | | TOTAL | |
|-------------------------------|----|--------|----|--------|----|---------|--|
| Surface area (ac) | | 1,323 | | 1,231 | | 2,554 | |
| Autosamplers (6) ¹ | \$ | 88,860 | \$ | 88,860 | \$ | 177,720 | |

Table 3. Estimated infrastructure costs for conducting field trials in the STA-1W Expansion Area during the Use of Soil Amendments/Management Techniques to Control P Flux Study.

¹Autosampler unit cost = \$29,620

Table 4. Estimated infrastructure and construction costs for a conceptual design of large-scale test facilities in select SAV cells of the STAs to be used in field trials during the Use of Soil Amendments/Management Techniques to Control P Flux Study.

| | STA-1E | STA-1W | STA-3/4 | STA-5/6 | TOTALS |
|-------------------------------------|------------------|-----------------|------------------|------------------|------------------|
| Candidate SAV treatment cells | 4N & 2 | 2B | 3B | 1B & 2B | - |
| Soil-amendment sub-cell (ac) | 318 | 48 | 325 | 612 | 1,303 |
| Limerock-cap sub-cell (ac) | 318 | 48 | 325 | 612 | 1,303 |
| Soil-Inversion sub-cell (ac) | 257 | 48 | 325 | 612 | 1,242 |
| New levees (mi) | 4 | 2 | 4 | 4 | 14 |
| Boat ramps (#) | 4 | 4 | 4 | 4 | 16 |
| Stage recorders (#) | 8 | 8 | 8 | 8 | 32 |
| Autosamplers (#) | 8 | 8 | 8 | 8 | 32 |
| Soil amendment (t/ac) | 3 | 3 | 3 | 3 | - |
| Limerock cap (cu. yd.) | 256,520 | 38,720 | 262,167 | 493,680 | 1,051,087 |
| DIRECT COSTS | | | | | |
| Land preparation ¹ | \$ 803,700 | \$ 129,600 | \$ 877,500 | \$ 1,652,400 | \$ 3,463,200 |
| New levee construction ² | \$ 1,791,440 | \$ 895,720 | \$ 1,791,440 | \$ 1,791,440 | \$ 6,270,040 |
| Boat ramps ³ | \$ 80,000 | \$ 80,000 | \$ 80,000 | \$ 80,000 | \$ 320,000 |
| Stage recorders ⁴ | \$ 140,000 | \$ 140,000 | \$ 140,000 | \$ 140,000 | \$ 560,000 |
| Autosamplers ⁵ | \$ 236,960 | \$ 236,960 | \$ 236,960 | \$ 236,960 | \$ 947,840 |
| Limerock cap ⁶ | \$ 5,899,960 | \$ 890,560 | \$ 6,029,833 | \$ 11,354,640 | \$ 24,174,993 |
| Soil amendment ⁷ | \$ 413,400 | \$ 62,400 | \$ 422,500 | \$ 795,600 | \$ 1,693,900 |
| Soil inversion ⁸ | \$ 257,000 | \$ 48,000 | \$ 325,000 | \$ 612,000 | \$ 1,242,000 |
| | \$ 9,622,460 | \$ 2,483,240 | \$ 9,903,233 | \$ 16,663,040 | \$ 38,671,974 |
| OVERHEAD COSTS ⁹ | | | | | |
| Mobilization/demob. (10.0%) | \$ 962,246 | \$ 248,324 | \$ 990,323 | \$ 1,666,304 | \$ 3,867,197 |
| Field office overhead (6.0%) | \$ 577,348 | \$ 148,994 | \$ 594,194 | \$ 999,782 | \$ 2,320,318 |
| Home office overhead (3.0%) | \$ 288,674 | \$ 74,497 | \$ 297,097 | \$ 499,891 | \$ 1,160,159 |
| OTHER COSTS ¹⁰ | | | | | |
| Sales tax (6.5%) | \$ 148,859 | \$ 38,416 | \$ 153,203 | \$ 257,777 | \$ 598,255 |
| Profit (6.0%) | \$ 687,044 | \$ 177,303 | \$ 707,091 | \$ 1,189,741 | \$ 2,761,179 |
| Bonds (1.5%) | \$ 184,299 | \$ 47,562 | \$ 189,677 | \$ 319,148 | \$ 740,686 |
| Contingency (15.0%) | \$ 1,842,995 | \$ 475,616 | \$ 1,896,771 | \$ 3,191,480 | \$ 7,406,862 |
| TOTAL COSTS | \$ 14,313,925 | \$ 3,693,953 | \$ 14,731,590 | \$ 24,787,165 | \$ 57,526,632 |
| Low Range -5% | \$ 13,598,229 | \$ 3,509,255 | \$ 13,995,011 | \$ 23,547,806 | \$ 54,650,300 |
| High Range +5% | \$ 15,029,621 | \$ 3,878,650 | \$ 15,468,170 | \$ 26,026,523 | \$ 60,402,964 |

¹Unit cost to clear and grub existing wetland vegetation in the three experimental sub-cells = $\$900/ac; {}^{2}$ Unit cost for construction of new levees = $\$447,860/mi; {}^{3}$ Boat ramp unit cost = $\$20,000; {}^{4}$ Stage recorder unit cost = $\$17,500; {}^{5}$ Autosampler unit cost = $\$29,620; {}^{6}$ Unit cost to purchase, transport to site and install a 6-inch layer of limerock over the entire surface area of the limerock-cap sub-cell = \$23/cu. yd.; 7 Unit cost to purchase and transport soil amendment to site = \$400/t + unit cost to apply soil amendment over entire surface area of the soil-amendment sub-cell = $\$100/ac; {}^{8}$ Unit cost to invert the surface layer of soil to a depth of 2 ft over the entire surface area of the soil-inversion sub-cell = $\$1,000/ac; {}^{9}$ Overhead costs calculated as a percentage of total direct costs; 10 Other costs calculated as a percentage of total direct costs + overhead costs.

| submerged aquatic vegetation (SAV). Cost estimates based of the submerged aquatic vegetation (SAV). | on applying | | |
|---|----------------|--|--|
| Total surface area dominated by SAV in the STAs $(ac)^1 =$ | 31,284 | | |
| SOIL AMENDMENT | | | |
| Mass of soil amendment applied over total SAV surface area (t) ² = | 93,851 | | |
| Cost to purchase, transport and apply soil amendment ³ = | \$ 40,668,680 | | |
| Cost of land preparation ⁴ = | \$ 28,155,240 | | |
| Overhead costs ⁵ = | \$ 13,076,545 | | |
| Other costs ⁶ = | \$ 17,209,077 | | |
| | \$ 99,109,542 | | |
| LIMEROCK CAP | | | |
| Volume of limerock applied over total SAV surface area (cu. yd.) ⁷ = | 25,235,437 | | |
| Cost to purchase, transport to site and install limerock cap ⁸ = | \$ 580,415,059 | | |
| Cost of land preparation ⁴ = | \$ 28,155,240 | | |
| Overhead costs ⁵ = | \$ 115,628,357 | | |
| Other costs ⁶ = | \$ 152,169,960 | | |
| | \$ 876,368,616 | | |
| SOIL INVERSION | | | |
| Cost of soil inversion over total SAV surface area ⁹ = | \$ 31,283,600 | | |
| Cost of land preparation ⁴ = | | | |
| Overhead costs ⁵ = | | | |
| Other costs ⁶ = | \$ 14,862,385 | | |
| | \$ 85,594,604 | | |

Table 5. Estimated costs to apply a soil amendment, install a limerock cap or invert the surface soil over the surface area of STA treatment cells dominated by

¹This area includes the entire surface area of treatment cells designated as "SAV cells" and the portion of the surface area of STA-2 Cells 2, 5 and 6 that is dominated by SAV; ²Mass calculation based on an application rate of 3 t/ac; ³Based on a unit cost of \$400/t to purchase and transport material to the site and \$100/ac to apply it; ⁴Based on a unit cost of \$900/ac to clear and grub existing wetland vegetation; ⁵Overhead costs calculated as a 19% markup of direct costs – see **Table 3** for details; ⁶Other costs calculated as a 25.0045% markup of direct costs – See Table 3 for details; ⁷Volume of limerock based on installing a 6-inch thick layer of material over the entire SAV surface area; ⁸Based on a unit cost of \$23/cu. yd. to purchase, transport to the site and install material; ⁹Based on a unit cost of \$1,000/ac to invert the surface layer of soil down to a depth of 2 ft.

Literature Cited

- Agyin-Birikorang, S., O. Oladeji, G. O'Connor, T. Obreza, and J. Capece. 2009. Efficacy of drinking-water treatment residual in controlling off-site phosphorus losses: a field study in Florida. Journal of Environmental Quality **38**:1076-1085.
- Ann, Y., K. Reddy, and J. Delfino. 2000. Influence of chemical amendments on phosphorus immobilization in soils from a constructed wetland. Ecological Engineering **14**:157-167.
- Arias, C. A., and H. Brix. 2004. Phosphorus removal in constructed wetlands: Can suitable alternative media be identified? Water Science and Technology **51**:267-274.
- Babin, J., E. Prepas, and Y. Zhang. 1992. Application of lime and alum to stormwater retention lakes to improve water quality. Water Pollution Research Journal of Canada **27**:365-381.
- Ballantine, D. J., and C. C. Tanner. 2010. Substrate and filter materials to enhance phosphorus removal in constructed wetlands treating diffuse farm runoff: A review. New Zealand Journal of Agricultural Research **53**:71-95.
- Berg, U., T. Neumann, D. Donnert, R. Nüesch, and D. Stüben. 2004. Sediment capping in eutrophic lakes–efficiency of undisturbed calcite barriers to immobilize phosphorus. Applied Geochemistry 19:1759-1771.
- Bottcher, D., T. DeBusk, H. Harper, G. O'Conner, and M. Wanielista. 2009. Technical Assistance for the Northern Everglades Chemical Treatment Pilot Project. SFWMD Project ID#: PS 1000093, Report prepared for the South Florida Water Management District, West Palm Beach, FL.
- Brooks, A. S., M. N. Rozenwald, L. D. Geohring, L. W. Lion, and T. S. Steenhuis. 2000. Phosphorus removal by wollastonite: A constructed wetland substrate. Ecological Engineering 15:121-132.
- Buda, A. R., G. F. Koopmans, R. B. Bryant, and W. J. Chardon. 2012. Emerging technologies for removing nonpoint phosphorus from surface water and groundwater: Introduction. Journal of Environmental Quality 41:621-627.
- Callahan, M. P., P. J. Kleinman, A. N. Sharpley, and W. L. Stout. 2002. Assessing the efficacy of alternative phosphorus sorbing soil amendments. Soil Science **167**:539-547.
- CH2M Hill. 2002. Periphyton-Based Stormwater Treatment Area (PSTA) Research and Demonstration Project: Soil Amendment Literature Review. Report prepared for the South Florida Water Management District, West Palm Beach, FL.
- CH2M Hill. 2003a. PSTA Research and Demonstration Project Field-Scale Soil Amendment Study Report. Report prepared for the South Florida Water Management District, West Palm Beach, FL.
- CH2M Hill. 2003b. PSTA Research and Demonstration Project Phase 1, 2, and 3 Summary Report. Report prepared for the South Florida Water Management District, West Palm Beach, FL.
- Chimney, M. J., O. Diaz, O. Villapando, and K. O'Dell. 2013. Alternative Treatment Technologies Evaluations - September 2011 to June 2013. Technical Publication WR-2013-3. South Florida Water Management District, West Palm Beach, FL. 156 pp.

- Chimney, M. J., Y. Wan, V. V. Matichenkov, and D. V. Calvert. 2007. Minimizing phosphorus release from newly flooded organic soils amended with calcium silicate slag: a pilot study. Wetlands Ecology and Management **15**:385-390.
- Christianson, L., A. Bhandari, and M. Helmers. 2009. Emerging technology: denitrification bioreactors for nitrate reduction in agricultural waters. Journal of Soil and Water Conservation **64**:139A-141A.
- Cooke, G. D., and R. H. Kennedy. 1981. Precipitation and inactivation of phosphorus as a lake restoration technique. EPA-600/3-81-012. Corvallis Environmental Research Laboratory, Office of Research and Development, US Environmental Protection Agency.
- Cooke, G. D., E. B. Welch, S. Peterson, and S. A. Nichols. 2005. Restoration and Management of Lakes and Reservoirs. CRC Press, Taylor & Francis Group, Boca Raton, FL.
- Cox, J. W., J. Varcoe, D. J. Chittleborough, and J. Van Leeuwen. 2005. Using gypsum to reduce phosphorus in runoff from subcatchments in South Australia. Journal of Environmental Quality **34**:2118-2128.
- Cucarella, V., and G. Renman. 2009. Phosphorus sorption capacity of filter materials used for on-site wastewater treatment determined in batch experiments—a comparative study. Journal of Environmental Quality **38**:381-392.
- Daniel, T., A. Sharpley, and J. Lemunyon. 1998. Agricultural phosphorus and eutrophication: A symposium overview. Journal of Environmental Quality **27**:251-257.
- DB Environmental Inc. 2013. The Use of Chemicals to Immobilize Phosphorus in Wetland Soils. PowerPoint presentation (unpublished).
- DeBusk, T. A., M. A. Langston, B. R. Schwegler, and S. Davidson. 1997. An evaluation of filter media for treating storm water runoff. Pages 82-89 *in* Proceedings of the Fifth Biennial Storm Water Research Conference.
- Ding, Y., B. Qin, H. Xu, B. Dong, and J. DBrookes. 2012. Comparison of efficacy of two Pinactivation agents on sediments from different regions of Lake Taihu: sediment core incubations. Fundamental and Applied Limnology/Archiv für Hydrobiologie **181**:271-281.
- Douglas, G., M. Robb, D. Coad, and P. Ford. 2004. A review of solid phase adsorbents for the removal of phosphorus from natural and wastewaters. Pages 291–320 *in* E. Valsami-Jones, editor. Phosphorus in Environmental Technologies: Principles and Applications. IWA Publishing, London.
- Drizo, A., C. Frost, J. Grace, and K. Smith. 1999. Physico-chemical screening of phosphateremoving substrates for use in constructed wetland systems. Water Research **33**:3595-3602.
- Elliott, H., G. O'Connor, P. Lu, and S. Brinton. 2002. Influence of water treatment residuals on phosphorus solubility and leaching. Journal of Environmental Quality **31**:1362-1369.
- Environmental Consulting & Technology, I. 2003. Copper Pilot Study, Prudential Property, Indian Summer Groves (No. KE 100-022) St. Lucie County, Florida. Report prepared for the South Florida Water Management District, West Palm Beach, FL.
- Environmental Consulting & Technology, I. 2006. Soil Inversion Demonstration Report For Picayune Strand Restoration Project Located Within Southern Golden Gates Estates

Collier County, Florida. Report prepared for the South Florida Water Management District, West Palm Beach, FL.

- Florida Department of Environmental Protection. 2011. List of Approved Flocculants.
- Gray, S., J. Kinross, P. Read, and A. Marland. 2000. The nutrient assimilative capacity of maerl as a substrate in constructed wetland systems for waste treatment. Water Research
 34:2183-2190.
- Hart, J. L. 2012. Evaluating the rates of nitrate removal for a nitrate containing, low organic carbon wastewater interacting with carbon-containing solid substrates. Master of Science thesis. Oregon State University, Corvallis, OR.
- Hickey, C. W., and M. M. Gibbs. 2009. Lake sediment phosphorus release management decision support and risk assessment framework. New Zealand Journal of Marine and Freshwater Research **43**:819-856.
- Hill, C. M., J. Duxbury, L. Geohring, and T. Peck. 2000. Designing constructed wetlands to remove phosphorus from barnyard runoff: a comparison of four alternative substrates. Journal of Environmental Science & Health Part A 35:1357-1375.
- Hoge, V. R., R. Conrow, M. Coveney, and J. Peterson. 2003. The application of alum residual as a phosphorus abatement tool within the Lake Apopka Restoration Area. Pages 1500-1513 *in* Proceedings of the Water Environment Federation, Baltimore, Md.
- Hopes, G. M. 2010. Hydraulic evaluation of wood chip media constructed wetlands for denitrification. Master of Science thesis. University of California, Davis, CA.
- Ippolito, J., K. Barbarick, and H. Elliott. 2011. Drinking water treatment residuals: A review of recent uses. Journal of Environmental Quality **40**:1-12.
- Jacobs, P. H., and U. Förstner. 1999. Concept of subaqueous capping of contaminated sediments with active barrier systems (ABS) using natural and modified zeolites. Water Research **33**:2083-2087.
- Johansson Westholm, L. 2006. Substrates for phosphorus removal—Potential benefits for onsite wastewater treatment? Water Research **40**:23-36.
- Jorge, J., J. Newman, M. J. Chimney, G. Goforth, T. Bechtel, G. Germain, M. Nungesset, D. Rumbold, J. Lopez, L. Fink, B. Gu, R. Berzotti, D. Campbell, C. Combs, K. Pietro, N. Iricanin, and R. Meeker. 2002. Chapter 4: Stormwater Treatment Areas and Advanced Treatment Technologies. 2002 Everglades Consolidated Report, South Florida Water Management District, West Palm Beach, FL.
- Kadlec, R. H., and S. Wallace. 2009. Treatment Wetlands, 2nd Edition. CRC Press, Taylor and Francis Group, Boca Raton, FL.
- Kirkkala, T., A.-M. Ventelä, and M. Tarvainen. 2012. Long-term field-scale experiment on using lime filters in an agricultural catchment. Journal of Environmental Quality **41**:410-419.
- Leader, J., E. Dunne, and K. Reddy. 2008. Phosphorus sorbing materials: sorption dynamics and physicochemical characteristics. Journal of Environmental Quality **37**:174-181.
- Lin, J., Y. Zhan, and Z. Zhu. 2011. Evaluation of sediment capping with active barrier systems (ABS) using calcite/zeolite mixtures to simultaneously manage phosphorus and ammonium release. Science of the Total Environment **409**:638-646.

- Lindstrom, S. M., and J. R. White. 2011. Reducing phosphorus flux from organic soils in surface flow treatment wetlands. Chemosphere **85**:625-629.
- Liu, G.-r., C.-s. Ye, J.-h. He, Q. Qian, and H. Jiang. 2009. Lake sediment treatment with aluminum, iron, calcium and nitrate additives to reduce phosphorus release. Journal of Zhejiang University SCIENCE A **10**:1367-1373.
- Makris, K. C., W. G. Harris, G. A. O'Connor, T. A. Obreza, and H. A. Elliott. 2005. Physicochemical properties related to long-term phosphorus retention by drinking-water treatment residuals. Environmental Science & Technology **39**:4280-4289.
- Malecki-Brown, L. M., and J. R. White. 2009. Effect of aluminum-containing amendments on phosphorus sequestration of wastewater treatment wetland soil. Soil Science Society of America Journal **73**:852-861.
- Malecki-Brown, L. M., J. R. White, and M. Sees. 2009. Alum application to improve water quality in a municipal wastewater treatment wetland. Journal of Environmental Quality **38**:814-821.
- Mara, D., M. Johnson, and M. A. Camerago-Velero. 2007. Rock Filters for Enhanced Phosphorus Removal. University of Leads.
- Meis, S. 2012. Investigating forced recovery from eutrophication in shallow lakes. Ph.D. Dissertation. Cardiff University, Cardiff, Wales, United Kingdom.
- Meis, S., B. M. Spears, S. C. Maberly, M. B. O'Malley, and R. G. Perkins. 2012. Sediment amendment with Phoslock[®] in Clatto Reservoir (Dundee, UK): Investigating changes in sediment elemental composition and phosphorus fractionation. Journal of Environmental Management **93**:185-193.
- Mortula, M., M. Gibbons, and G. A. Gagnon. 2007. Phosphorus adsorption by naturallyoccurring materials and industrial by-products. Journal of Environmental Engineering and Science **6**:157-164.
- Novak, J., and D. Watts. 2005. An alum-based water treatment residual can reduce extractable phosphorus concentrations in three phosphorus-enriched coastal plain soils. Journal of Environmental Quality **34**:1820-1827.
- O'Connor, G., S. Brinton, and M. Silveira. 2005. Evaluation and selection of soil amendments for field testing to reduce P losses. Soil and Crop Science Society of Florida, Proceedings **64**.
- Pekrun, C., H.-P. Kaul, and W. Claupein. 2003. Soil Tillage for Sustainable Nutrient Management. Pages 83-113 *in* A. El Titi, editor. Soil tillage in agroecosystems. CRC Press LLC, Boca Raton, FL.
- Penn, C. J., R. B. Bryant, P. J. Kleinman, and A. L. Allen. 2007. Removing dissolved phosphorus from drainage ditch water with phosphorus sorbing materials. Journal of Soil and Water Conservation **62**:269-276.
- Reddy, K., J. White, M. Fisher, H. Pant, Y. Wang, K. Grace, and W. Harris. 2002. Potential impacts of sediment dredging on internal phosphorus loading in Lake Okeechobee. Summary report. Report prepared for South Florida Water Management District, West Palm Beach, FL.

- Reddy, K. R., G. Bao, O. G. Olila, and D. L. Stites. 1998. Effects of Chemical Amendments on Marsh Soil Chemistry and Nutrient Flux. First Draft, Final Report Contract No. 96W258, University of Florida & St Johns River Water Management District, Report submitted to St Johns River Water Management District, Palatka, FL.
- Reddy, K. R., and R. D. DeLaune. 2008. Biogeochemistry of Wetlands: Science and Applications. CRC Press, Taylor & Frances Group, Boca Raton, FL.
- Sanville, W. D., A. R. Gahler, J. A. Searcy, and C. F. Powers. 1976. Studies on lake restoration by phosphorus inactivation. EPA-600/3-76-041, US Environmental Protection Agency, Office of Research and Development, Corvallis Environmental Research Laboratory.
- Schipper, L. A., W. D. Robertson, A. J. Gold, D. B. Jaynes, and S. C. Cameron. 2010. Denitrifying bioreactors—An approach for reducing nitrate loads to receiving waters. Ecological Engineering 36:1532-1543.
- Shaw Environmental, I. 2007. Work Order #8 Report Soil Inversion Pilot Test Agler Property, St. Lucie County, Florida. Report prepared for the South Florida Water Management District, West Palm Beach, FL.
- South Florida Water Management District. 2009. Influence of Tilling on Phosphorus Sequestration from Littoral Zone Sediments of Lake Okeechobee, Florida. South Florida Water Managment District, West Palm Beach, FL.
- South Florida Water Management District, and URS. 2014. Istokpoga Marsh Water Improvement District Water Quality Improvement Project Copper Corrective Action Soil Inversion Demonstration Plan.
- Spears, B. M., S. Meis, A. Anderson, and M. Kellou. 2013. Comparison of phosphorus (P) removal properties of materials proposed for the control of sediment P release in UK lakes. Science of the Total Environment **442**:103-110.
- Strang, T., and D. Wareham. 2006. Phosphorus removal in a waste-stabilization pond containing limestone rock filters. Journal of Environmental Engineering and Science **5**:447-457.
- Struve, D., and M. Zhou. undated. The Use of dinoSoil and Other Materials for Surface Water Reduction and Manure P Stabilization. Unpublished Report, South Florida Water Management District, West Palm Beach, FL.
- Vohla, C., M. Kõiv, H. J. Bavor, F. Chazarenc, and Ü. Mander. 2011. Filter materials for phosphorus removal from wastewater in treatment wetlands—a review. Ecological Engineering **37**:70-89.
- Water and Soil Solutions, L. 2009. Tilling Practices for Phosphorus, Sediment and Vegetation Management in Lake Okeechobee Phase I. Report prepared for South Florida Water Management District, West Palm Beach, FL.
- Welch, E. B., and G. D. Cooke. 1999. Effectiveness and longevity of phosphorus inactivation with alum. Lake and Reservoir Management **15**:5-27.
- Welch, E. B., C. L. DeGasperi, D. E. Spyridakis, and T. J. Belnick. 1988. Internal phosphorus loading and alum effectiveness in shallow lakes. Lake and Reservoir Management 4:27-33.

- WRScompass, I. 2009. United States Sugar Corporation Sugar Cane Fields Soil Inversion Pilot Study. Report prepared for the South Florida Water Management District, West Palm Beach, FL.
- Xu, D., S. Ding, Q. Sun, J. Zhong, W. Wu, and F. Jia. 2012. Evaluation of in situ capping with clean soils to control phosphate release from sediments. Science of the Total Environment 438:334-341.
- Yamada, H., M. Kayama, K. Saito, and M. Hara. 1987. Suppression of phosphate liberation from sediment by using iron slag. Water Research **21**:325-333.
- Ye, H.-P., F.-Z. Chen, Y.-Q. Sheng, G.-Y. Sheng, and J.-M. Fu. 2006. Suppression of phosphate liberation from eutrophic lake sediment by using fly ash and ordinary Portland cement. Journal of Environmental Science and Health Part A **41**:1655-1666.
- Zamorano, M., H. Zhao, K. Grace, M. Jerauld, T. A. DeBusk, T. Piccone, and D. Ivanoff. 2014. STA-3/4 Periphyton-based Stormwater Treatment Area Project. Pages 5B-59 - 80 in M. J. Chimney, editor. Chapter 5B: Performance of the Everglades Stormwater Treatment Areas, 2014 South Florida Environmental Report, South Florida Water Management District, West Palm Beach, FL.
- Zvomuya, F., C. J. Rosen, and S. C. Gupta. 2006. Phosphorus sequestration by chemical amendments to reduce leaching from wastewater applications. Journal of Environmental Quality **35**:207-215.

Appendix 1 – Revised Conceptual Design for Conducting Field Trials in the STAs as Part of the Use of Soil Amendments/Management Techniques to Control P Flux Study

Based on review comments received from the Technical Representatives on the Detailed Study Plan (DSP)⁸ and additional direction provided by District management, the conceptual design for the large-scale test facilities (LSTFs) in the STAs has been revised and we now propose two different options for conducting field trials. The first option is to use Cells 7 and 8 of the STA-1W Expansion Area (EA) to investigate the efficacy of soil inversion⁹. The second option is to partition several existing SAV treatment cells into a number of "sub-cells" and conduct 6field trials of different technologies within these experimental units.

Field-trial Option #1 - STA-1W Expansion Area

The STA-1W EA will consist of three new cells, Cell 6, 7 and 8, located immediately west of the existing STA-1W facility (**Appendix Fig. 1**). All the soil in Cell 7 will be inverted to a depth of 2 ft while all the soil in Cell 8 will be left undisturbed¹⁰. The treatment performance of Cell 7 (1,323 ac, the experimental unit) will be compared to that of Cell 8 (1,231 ac, the control) to evaluate the efficacy of soil inversion to reduce outflow TP concentrations. The only new infrastructure that may be needed for this option is to install autosamplers at the inflow and outflow water-control structures in Cells 7 and 8 if autosamplers at these locations are not already part of the STA-1W EA design. We do not anticipate that any other modifications will be required to utilize these cells.

The current construction schedule calls for all STA-1W EA cells to be flow-capable by December 31, 2018. It is estimated that it will take 12 to 18 months for the aquatic vegetation community to become established after the facility is flooded and the cells to meet their phosphorus start-up criterion. Data collection can begin as soon as flow-through operation in the STA-1W EA is permitted, which may be as soon as January 2020 and will last 4 to 5 years.

Field-trial Option #2 - LSTFs in existing SAV cells

Treatment cells to be employed in this option were selected based on their configuration, which allows for the construction of a number of rectangular sub-cells arranged in parallel within each STA; candidate treatment cells include STA-1E Cells 4N and 2, STA-1W Cell 2B, STA-3/4

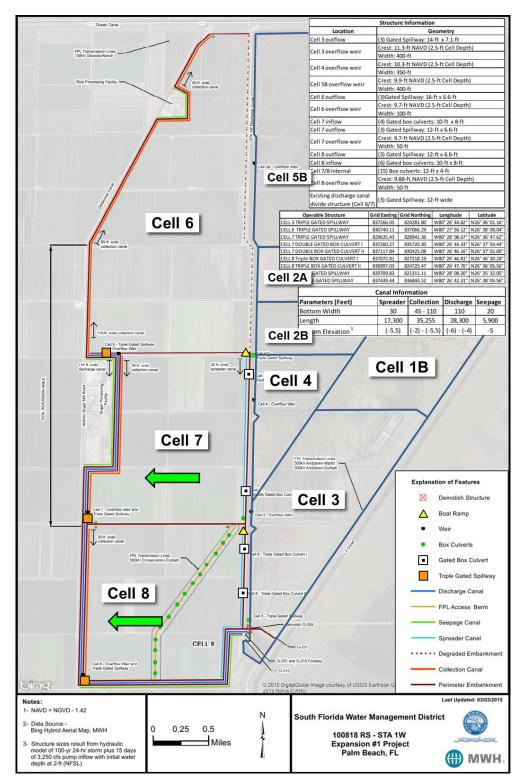
⁸ The Technical Representatives expressed concerns that the 10-ac experimental cells originally proposed for the field trials in the DSP would have performance artifacts due to their relatively small size and not be representative of the full-scale STAs. There also were issues related to achieving desired hydraulic retention times and how to mimic the hydrology of the STAs, especially peak flows during storm events, in the 10-ac experimental cells.

⁹ Surface soil in portions of the STA-1W Expansion Area will be inverted during construction to bury its high copper concentrations. We would take advantage of this situation to investigate whether soil inversion can reduce outflow TP concentrations.

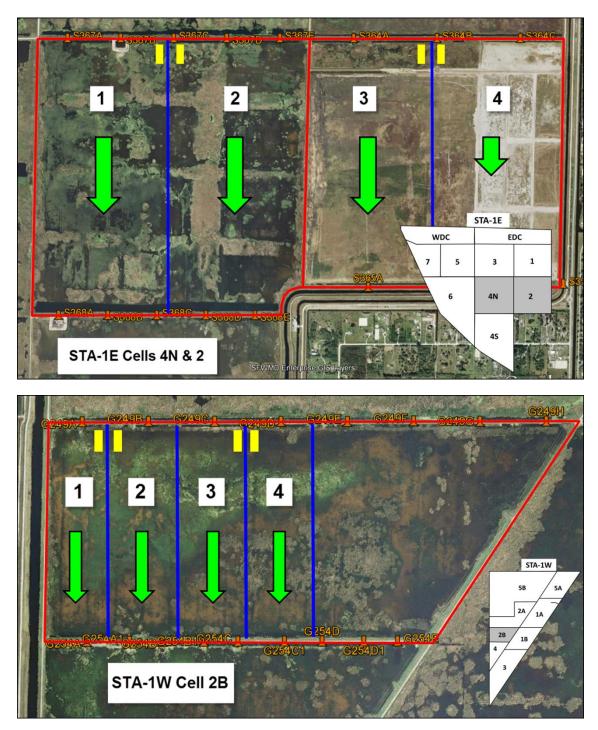
¹⁰ Approximately one-half of the soil in Cell 6 will be inverted while the remaining soil will be left undisturbed. Water that flows over each of these areas will mix within the cell. Because these waters cannot be sampled separately, Cell 6 is not suitable for conducting an experiment on the influence of soil inversion on treatment performance.

Cell 3B and STA-5/6 Cells 5-1B and 5-2B (**Appendix Fig. 2**). The plant community in all these cells is dominated by SAV. This new design has much larger experimental units (the sub-cells range in size from 48 to 612 ac, depending on the STA) and utilizes existing water control structures for water delivery rather than relying on small pumps to manage inflow and outflow as originally proposed in the DSP. New levees will be constructed that together with existing levees will create four sub-cells within each of the above-mentioned treatment cells. Each sub-cell will have dedicated head- and tailwater stage recorders and inflow and outflow autosamplers. Our intent is to operate the candidate treatment cells the same as the other treatment cells in each respective STA during the study and to apply the same hydraulic load to each sub-cell within a treatment cell. One sub-cell will be used as a control, i.e., it will receive no experimental manipulation, while the other sub-cells will be manipulated. The sub-cells will be cleared of all wetland vegetation before treatments are applied and the aquatic plant communities reestablished before conducting the experiment.

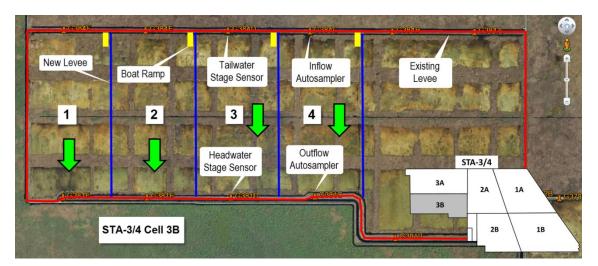
It is estimated that the design and construction of the new levees plus application of the experimental treatments within the sub-cells (a soil amendment, a limerock cap and soil inversion) will require 18 to 24 months followed by an additional 6 to 12 months for aquatic vegetation to grow in. Data collection could start 24 to 36 months after a GO decision is made to proceed with Phase III of the study and will continue for 4 to 5 years.

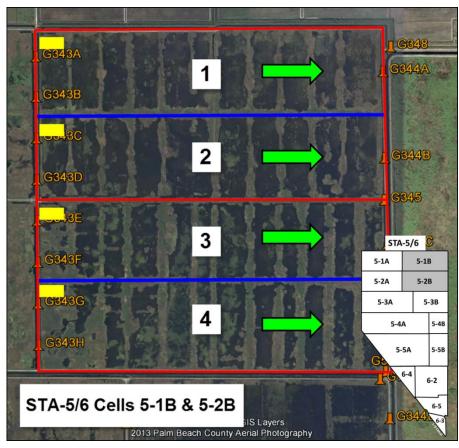


Appendix Figure 1. Schematic of the STA-1W Expansion Area showing Cells 7 and 8 that are proposed for use as large-scale test facilities in the Use of Soil Amendments/Management Techniques to Control P Flux Study. Yellow triangles indicate existing boat ramps; white and orange squares indicate existing inflow and outflow water control structures, respectively; green arrows indicate the direction of flow in Cells 7 and 8.



Appendix Figure 2. Schematics of proposed large-scale test facilities in candidate STA Treatment Cells (STA-E Cells 4N and 2, STA-1W Cell 2B, STA-3/4 Cell 3B and STA-5/6 Cells 5B-1B and 5B-2B) for the Use of Soil Amendments/Management Techniques to Control P Flux Study. Red lines indicate existing levees; blue lines indicate new levees to be constructed; yellow rectangles represent new boat ramps; orange symbols show locations of existing water control structures; green arrows indicate the direction of flow. Numbers identify the new sub-cells within each treatment cell. Map inserts show the location of the candidate treatment cell (shaded gray) within its respective STA.





Appendix Figure 2. (Continued).